ce & lechnolog

Impacts of Land-Use Changes on the Lakes across the Yangtze Floodplain in China

Cong Xie,[‡] Xin Huang,^{*,†,‡} Hongqiang Mu,[§] and Wei Yin^{||}

[†]School of Remote Sensing and Information Engineering and [‡]State Key Laboratory of Information Engineering in Surveying, Mapping and Remote Sensing, Wuhan University, Wuhan 430079, People's Republic of China

[§]Yangtze Valley Water Resources Protection Bureau, Changjiang Water Resources Commission, Wuhan 430010, People's Republic of China

Yangtze River Water Resources Protection Science Institute, Wuhan 430051, People's Republic of China

Supporting Information

ABSTRACT: The middle and lower Yangtze (MLY) floodplain has one of the most densely distributed lake clusters in China but suffered from long-term lake reclamation and wetland degradation due to intensive cultivation, fish rearing, and urban expansion over the past several decades. As a land-use alternation to support human life, the conversion of lakes to cropland, aquaculture ponds, and human settlements provides essential ecosystem goods at the expense of the deterioration of wetland environment. To quantify the driving factors of lake changes, we investigated the land-use transitions from lakes ($\geq 1 \text{ km}^2$) between 1975 and 2015 using Landsat remote sensing data. We found that the dramatic decline in lake area (a net decrease of $13.8 \pm 1.4\%$) over the four decades was largely attributed to human-induced transformation from lakes to cropland, fish ponds, and builtup areas, accounting for 34.6%, 24.2%, and 2.5% of the total area reduction, respectively. The remaining loss, associated with vegetation (37.3%) and bare land (1.4%) and coming mainly from China's two largest freshwater lakes (Poyang and Dongting), can be explained by climate variation, sediment deposition, and hydrological regulation. These findings shed



new light on the quantitative impacts of human activities and climate variation on lake changes and provide a scientific foundation for wetland management decision-making.

INTRODUCTION

Lakes, acting as a regulator of the carbon cycle and global climate in the terrestrial system, constitute a significant component of global water resources, providing an irreplaceable environment for aquatic fauna and flora as well as for human survival.¹ However, due to intensive human activities and climate change, lakes all over the world have undergone drastic changes,^{2–7} drawing considerable attention to climate change,^{8,9} water resources,^{10,11} and environmental pollution.¹² Lake shrinkage and wetland degradation at the global scale have caused extensive concern about how to accurately track lake changes and driving factors, as well as to ease their impact on ecological systems.

The middle and lower Yangtze (MLY) floodplain is covered with the largest freshwater lake cluster in East Asia, which plays a critical role in flood control, runoff regulation, human consumption, and social development. As an important commodity grain base, accounting for 24% of China's agricultural production, the MLY floodplain is one of the most densely populated regions, sustaining 25% of the total population (338 million) of China.¹³ Due to intensive cultivation, fish rearing, and urban expansion over the past half century, the region is now considered to be one of the most fragile ecological systems in the world¹⁴ and has experienced rapid shrinkage of the lake area,¹⁵⁻¹⁸ severe degradation of the wetland ecosystem,¹⁹ deterioration of water quality,²⁰ and decreased biodiversity.²¹

Although a few studies have focused on either individual lake regions (e.g., Poyang, Dongting, Tai, and Chao), $^{22-25}$ or lake groups in the MLY floodplain, $^{15-17,26}$ the quantitative impacts of climatic and anthropogenic factors on the lake changes across the entire floodplain still remain unclear. For example, the freshwater lakes in the Central Yangtze area of China have undergone dramatic changes over the past several decades, mainly due to intensive land reclamation, resulting in serious negative ecological consequences.^{17,18} Some studies have also reported that the changes of lakes in the MLY floodplain were more attributed to human activities (e.g., impoldering practices, enclosures, and urban expansion) than natural factors (e.g., climate change and sediment accumulation).^{5,15,27} The previous understanding of the various potential driving forces behind lake changes is mainly based on qualitative description or correlation analysis, yet quantifying such drivers of lake degradation is of great importance to the lake regions suffering from both climate change and anthropogenic activities.

Received:	August 22, 2016
Revised :	March 9, 2017
Accepted:	March 12, 2017
Published:	March 13, 2017

Downloaded via WUHAN UNIV on December 2, 2019 at 12:02:56 (UTC). See https://pubs.acs.org/sharingguidelines for options on how to legitimately share published articles.



Figure 1. Distribution of lake groups in the middle and lower reaches of the Yangtze River. The MLY floodplain consists of the most alluvial regions in the middle reach (e.g., Dongting, Jianghan, and Poyang plains), and the lower reach (e.g., Wanzhong and Yangtze Delta plains) of the Yangtze Basin. The inset shows the location of the study area in China.

Therefore, the objective of this study is to quantify the lake changes and associated driving forces across the MLY floodplain by exploiting four decades of Landsat remotesensing images. Furthermore, we also discussed the negative ecological impacts (e.g., flood disasters, biodiversity loss, and water quality) of lake degradation caused by excessive human exploitation. The long-term quantitative results provide a scientific basis for balancing the inherent trade-offs between human consumption and lake protection and help inform future decision-making for wetland management and ecosystem restoration.

MATERIALS AND METHODS

Study Area. The middle and lower Yangtze (MLY) basin, situated in eastern China, refers to the section extending from Three Gorges Dam (TGD) to the mouth of the Yangtze River, with an area of $7.8 \times 10^5 \text{ km}^2 (106^{\circ}7' - 121^{\circ}47' \text{ E}, 24^{\circ}30' - 121^{\circ}47' \text{ E})$ 33°54′ N) and an elevation of less than 50 m (Figure 1). Influenced by the subtropical monsoon climate, this region is characterized by abundant rainfall, with an annual mean precipitation (AMP) of 1000-1400 mm and an annual mean temperature (AMT) of 14-18 °C. The MLY basin is wellknown for the most dense group of freshwater lakes in East Asia, representing 38.17% of the total freshwater area in China.¹⁵ Containing over 75% of the total surface area of the lakes in the MLY basin, the MLY floodplain consists of a series of alluvial plains, which can be divided into five lake subregions (i.e., Dongting, Jianghan, Poyang, Wanzhong, and the Yangtze Delta; Figure 1) according to the spatial distribution of lakes and geographical differences.²⁶ A total of four of the five largest freshwater lakes in China, i.e., Poyang, Dongting, Tai, and

Chao, constitute vitally important components of the lake systems in the MLY floodplain, providing an indispensable ecological environment for aquatic fauna and flora. Many wetlands in the floodplain serve as internationally important habitats for endangered species and migratory waterfowl, and nine sites have been designated as Ramsar sites²⁸ (Table S1). However, the lake resources of the floodplain have suffered from long-term extensive exploitation to support rapid population growth and economic development, leading to the increasingly irreconcilable conflict between human consumption and lake conservation. Consequently, the regional ecosystem has experienced a serious degradation over the past few decades, with a significant vulnerability to human activities and climate change. For example, the surface area of Dongting Lake decreased by 49.2% between the 1930s and 1998, mainly due to intensive impoldering practices.¹⁸

Data. In this study, remotely sensed data were obtained from the United States Geological Survey (USGS; http://www. usgs.gov/), including Landsat Multispectral Scanner (MSS) images taken around 1975, Thematic Mapper (TM), and Enhanced Thematic Mapper Plus (ETM+) images taken around 1990, 2000, and 2010, and Operational Land Imager (OLI) images taken around 2015 (Table S2). Cloud-free images were acquired in the wet season to avoid the influence of seasonal variation on the estimation of the lake surface area. By the comparison of multitemporal images acquired over at least two years for each study period, the suitable images (excluding anomalous rainfall or drought conditions) were chosen so that the impacts of particularly wet and dry years did not skew the results (e.g., lakes area as well as land use changes). Additionally, the monthly air temperature and precipitation data from 1975 to 2015 from 24 meteorological stations over the MLY floodplain were acquired from the China Meteorological Data Sharing Service System (http://cdc.cma. gov.cn/). Agricultural data, aquaculture area, human population, and urban area were documented from the local statistical year book^{29–33} and were used to provide auxiliary materials to analyze the effects of human activities on lake changes.

Methods. All the remote sensing images were well georeferenced based on topographical maps, and the MSS images were resampled to the resolution of TM images (30 m). Normalized difference water index (NDWI³⁴ and MNDWI,³⁵ Table 1) was calculated to delineate the water surface area of

Table 1. Information Indexes Derived from Band Operation of Landsat $Images^a$

information indexes	abbreviation	formula
normalized difference water index	NDWI ^b	$NDWI = \frac{\text{green} - \text{NIR}}{\text{green} + \text{NIR}}$
modified normalized difference water index	MNDWI ^c	$MNDWI = \frac{green - SWIR}{green + SWIR}$
normalized difference vegetation index	NDVI	$NDVI = \frac{NIR - red}{NIR + red}$
normalized difference built-up index	NDBI	$NDBI = \frac{SWIR - NIR}{SWIR + NIR}$
bare soil index	BSI	$BSI = \frac{(SWIR + red) - (NIR + blue)}{(SWIR + red) + (NIR + blue)}$

^aSpectral bands: near-infrared (NIR) and short-wave infrared (SWIR) bands. ^bNDWI is calculated for Landsat MSS images. ^cMNDWI is calculated for Landsat TM/ETM+/OLI images.

lakes ($\geq 1 \text{ km}^2$) for each selected image. Note that in this study, we only focused on natural lakes, and hence, rivers, reservoirs, and artificial ponds were removed. To strictly control the accuracy of lake area extraction, we manually delineated the water boundary of each lake by visual interpretation of images. The software ArcGIS 10.3 was used to calculate the water surface area of each lake. Following the careful delineation of lake surface area, the lake database of the MLY floodplain was established.

Changes in lake area (i.e., shrinkage, expansion, and no change) were further detected by overlapping the lake maps from different periods of 1975-2015 in the software ArcGIS 10.3. The change rate in surface water area was calculated as the ratio of the difference of the lake area between two periods (A_{t1}) and A_{t2} to the lake area in the earlier period, i.e., $(A_{t2} - A_{t1})/(A_{t2} - A_{t1})/(A_{t$ $A_{t1} \times 100\%$, and thus, a change rate of -100% suggests that the lakes had vanished. Furthermore, an elaborate classification scheme combining several information indexes (i.e., modified normalized difference water index (MNDWI),35 normalized difference vegetation index (NDVI),³⁶ normalized difference built-up index (NDBI),³⁷ and bare soil index (BSI),³⁸ Table 1) was employed to quantitatively identify the land-use transitions from lakes to others, including aquaculture ponds, vegetation, built-up areas, bare land, and agricultural land. The empirical threshold values for the information indexes were manually selected for image classification based on the ENVI 5.1 software. Because cropland and vegetation are spectrally similar in summer images, we collected spring images to distinguish cropland from vegetation.³⁹ An example of Sha Lake (a typical lake suffering from serious land reclamation in Wuhan) is shown in the Abstract graphic. To ensure the reliability of the quantitative results, we validated carefully the detected land-use

changes of each lake by visual examination referring to highresolution images from Google Earth.⁴⁰ In this study, the uncertainties were mainly associated with the 40 year remote sensing data, the image-processing methods, and the census data from local statistical yearbooks, which were presented in Text S1.

RESULTS AND DISCUSSION

Changes in Lake Surface Area between 1975 and 2015. The area and number of all the lakes across the MLY floodplain (Table S3) were derived from Landsat images from 1975 to 2015. To facilitate an analysis of different-sized lakes, the lakes were categorized into four groups: 1-10, 10-50, 50-100, and >100 km².⁴¹ A total of 389 lakes (\geq 1 km²) were extracted for 1975, of which the four groups accounted for 78.1%, 13.6%, 4.2% and 4.1%, respectively. Over the past four decades, the total area of lakes has decreased by 2132.3 ± 219.6 km^2 (13.8 ± 1.4%), and the number of lakes experienced a net reduction of 26 (from 389 to 363), including 37 vanished lakes and 11 newly formed lakes (Table S3). The lakes in Dongting Plain experienced the most-serious decline (855.1 \pm 131.8 km²) in surface area while, relatively, the slightest decrease of lake area $(153.7 \pm 27.8 \text{ km}^2)$ occurred in the Yangtze Delta. An increase in the area and number of lakes can be observed in the size group 10-50 km² (Table S3), on account of the degradation of the lakes from the size group $50-100 \text{ km}^2$. The water surface area of most lakes (61.6%) experienced a significant reduction between 1975 and 2015, the majority of which showed a change rate of -50-0% (Figure 2).



Figure 2. Comparison of lake area between 1975 and 2015 for each subregion. The inset shows the frequency distributions of change rate in lake area, suggesting that most lakes shrunk with a change rate of -50-0% and a few lakes vanished (change rate of -100%). The solid and dashed lines represent the fitted and 1:1 lines, respectively.

To investigate the temporal changes of lakes, 85 medium and large lakes ($\geq 10 \text{ km}^2$) were further studied over the five nominal periods: 1975, 1990, 2000, 2010, and 2015. The total surface area of the lakes across the entire floodplain showed a rapid reduction during 1975–2000, with a slight increase after 2000 (Figure 3). To describe the change trend of the lake area on a statistical basis, the paired samples *t*-test was employed using the Statistical Package for the Social Sciences (SPSS) software to compare the area of each individual lake between different periods (e.g., 1975 versus 1990 and 1990 versus 2000)



Figure 3. Changes of lake area in each subregion during the five periods: 1975, 1990, 2000, 2010, and 2015. The error bars are taken as the uncertainties of lake surface areas. The paired samples *t*-test was employed to compare the lake area between different time periods. The single asterisk (*) in blue suggests significant decrease in lake area at 80% confidence level, double asterisks in blue indicate decrease at 95% confidence level, and the red asterisk denotes increase in lake area at 80% confidence level.

in all subregions (Figure 3). Statistical significance was determined at the 80% or 95% confidence level. The lakes in Jianghan Plain shrunk notably (p < 0.2) during 1975–2000, whereas the water surface area showed a significant increase (p < 0.2) in this region from 2000 to 2010. Wanzhong Plain and the Yangtze Delta witnessed a remarkable reduction (p < 0.05 or p < 0.2) of lake area in the period of 1975–2010, while a statistically insignificant decrease (p > 0.2) occurred in Dongting Plain and Poyang Plain.

Quantitative Analysis of Driving Forces of Lake Changes. To identify the quantitative impacts of natural and anthropogenic factors, lake changes (transition to aquaculture ponds, vegetation, developed areas, bare land, and cropland) across the whole floodplain were obtained by accurate interpretation of Landsat data over the past four decades (Figure 4a). The results revealed that the significant shrinkage of lakes was predominately driven by human activities, i.e., lakes were directly converted into cropland, aquaculture ponds, and built-up areas (see examples in Figure S1a-c), accounting for 34.6% (898.3 km²), 24.2% (627.7 km²), and 2.5% (64.5 km²) of



Figure 4. Land-use changes from lakes $(\ge 1 \text{ km}^2)$ to other land-use types (including aquaculture ponds, vegetation, built-up areas, bare land, and agricultural land) in the MLY floodplain during 1975–2015. (a) Stacked charts represent the area of each land-use type to which the lakes were converted. (b) Stacked charts indicate the number of lakes in each "major conversion type" group.

the total loss in lake area, respectively. Furthermore, the conversion of lakes into vegetation (Figure S1d) and bare land can explain the remaining reduction in surface area (i.e., 967.0 and 35.5 km² with 37.3% and 1.4% in proportion, respectively), as a result of local climate variation, sediment siltation, and hydrological regulation. Furthermore, we identified the major land-use type for each lake in terms of the area of change, hereafter referred to as the "major conversion type". In this way, the lakes were categorized into five groups according to the "major conversion type", and the number of lakes in each group was derived for all the subregions (Figure 4b). Aquaculture pond or cropland was the "major conversion type" for most lakes, 50.1% and 37.3% in number, respectively. The percentage of vegetation as the "major conversion type" (7.5%) showed a significant difference compared to that of vegetation in area (37.3%, as shown in Figure 4a), which suggested a large change to vegetation occurred in a few lakes.

Human-Induced Land-use Changes of Lakes. Lake changes exhibited different spatial patterns from west to east along the Yangtze River, depending on the regional ecological environment and economic conditions. The lakes in Jianghan Plain and the Yangtze Delta were primarily converted into aquaculture ponds, while Wanzhong Plain was dominated by the transition to cropland. The abundant freshwater resources in Jianghan Plain facilitated the rapid expansion of aquaculture ponds in Hubei province (Figure 5c) but at the cost of 60.3% $(260.4 \text{ km}^2 \text{ in area})$ loss in lakes devoted to aquatic breeding. For example, Hong Lake, the largest lake in Jianghan Plain, had a decrease of 30.4 km^2 (98.7% of the aggregate reduction) as a result of enclosure practices for fish farming. As an important economic activity in the floodplain, fish rearing has provided locals with a plentiful aquatic resource, which has gradually encroached on the natural lakes through the construction of barriers for growing aquaculture area (Figure 5c).²⁷ Along with the rapid development of fishery industry, the overexploitation of fish resources has posed a serious threat to the survival of native fauna and flora, resulting in severe decline of biodiversity.²¹ The lakes in the Yangtze Delta showed a similar pattern, i.e., a large amount of lakes were turned into fish ponds, leading to a reduction of 112.6 km², induced by the increasing market demands for aquatic products. For instance, Tai Lake, China's third-largest freshwater lake, supporting one of the most important fisheries for crabs, carp, and eels, suffered from the over-exploitation of freshwater resources and the deterioration of water quality.²⁵ In Wanzhong Plain, extensive impoldering for food production to support a growing population in Anhui Province (Figure 5e) was the leading driving force, inducing 59.5% (291.2 km² in area) of the total loss in lakes. Due to intensive land reclamation and overuse of chemical fertilizers, Shengjin Lake national nature reserve (Table S1) experienced wetland degradation and environmental pollution from 1986 to 2002.⁴² Overall, human-induced land-use transitions from lakes to cropland, fish ponds, and urban areas represent the major drivers of lake shrinkage in the above three plains (i.e., Jianghan, Wanzhong, and Yangtze Delta), accounting for 84.6% of the total loss in lake area. However, natural factors such as meteorological variation, soil erosion, and sediment accumulation, inducing lake changes to vegetation and bare land, can explain the additional 15.4% of area reduction. Over the past 40 years, a consistent increase (p < 0.001) has been observed in interannual temperature (Figure 5a), whereas no significant trend (p > 0.1) can be seen in interannual precipitation change (Figure 5b). Therefore, it was



Figure 5. Changes in climate and human activities in the MLY floodplain over the past few decades. (a, b) The annual mean temperature (AMT) anomaly and annual mean precipitation (AMP) anomaly in each subregion, respectively. (c-f) The aquaculture area, cultivated area, human population, and developed area in each subregion, respectively.

believed that the climate change may not be considered as a major driving factor for lake shrinkage in the MLY floodplain.^{15,18} Compared to climate change, soil erosion has a more significant impact on lake changes, which can increase the influx of sediment into lakes and raise the lake bed levels.^{15,27} Although the free connection between the Yangtze River and most lakes in the MLY floodplain has been blocked as a result of local floodgate control or channel diversion,^{17,26} these lakes have experienced accelerated sediment accumulation due to soil erosion from nearby cultivated lands.²⁷ For example, Hong Lake suffered from severe soil erosion and sedimentation from 1976 to 1983, inducing a rapid expansion of hydrophytes at an annual average speed of 7.41 km².⁴³

Lake Degradation in Poyang and Dongting Plains. Poyang and Dongting plains witnessed the largest degradation from lakes to vegetation, accounting for 32.4% (188.4 km²) and 69.4% (631.0 km²) of the reduction of lake area in the respective regions (Figure 4a), mainly from China's two largest freshwater lakes (Poyang and Dongting). Connected freely to the Yangtze River, the two lakes experienced a dramatic reduction in surface area over the past few decades, leading to widespread exposure of wetland vegetation (187.2 and 620.1 km² in area, respectively).^{23,44–46} Long-term sediment siltation due to a large amount of water-soil exchange with the Yangtze River was an important factor, resulting in the loss of the water storage capacity of lakes.¹⁵ For instance, Dongting Lake had an annual sediment input of 1.9×10^8 t, mainly from the Yangtze River, leading to a high sedimentation rate of 10-30 mm/ year.²⁷ Furthermore, TGD, the world's largest hydroelectric dam, with a reservoir storage capacity of >39 km^{3,47} aggravated

the water-level decline of the Yangtze downstream lake systems (e.g., Poyang and Dongting), accelerating the emergence of wetland vegetation. 44

In addition to the transition to wetland vegetation, Poyang and Dong plains have also suffered from excessive human utilization (i.e., transformation from lakes into cropland, fish ponds, and built-up areas), representing 66.8% and 29.6% of the area reduction in the respective regions. For example, extensive land reclamation has been another dominant driver for the shrinkage of Poyang Lake, resulting in a decrease of 249.8 km² in surface area for agricultural use. Dongting Lake has also undergone a sharp shrinkage from 4955 km² in the 1930s to \sim 2266 km² in the early 1990s as a result of the practice of intensive impoldering.¹⁸ Generally, the quantitative results have demonstrated that in Central Yangtze (including Dongting, Poyang, and Jianghan plains), human-induced lake conversions were the major driving factors (53.5%) for the decline in the lake surface area, while natural factors (e.g., climate change and sediment siltation) and hydrological regulation (e.g., TGD) should be responsible for the remaining loss (46.5%), which triggered the transformation from lakes to vegetation and bare land. For the entire MLY floodplain, when Dongting and Poyang Lakes are excluded, the direct land-use transitions from lakes to fish ponds, cropland, and urban areas can explain 86.1% of the lake area loss, indicating local anthropogenic activities were the dominant driver of lake shrinkage in the entire study area.

Relative Contributions of Anthropogenic and Natural Drivers. To further identify the relative contributions of human activities and natural factors on the lake changes, we



Figure 6. Boxplots show the percentage distributions of land use (including aquaculture pond, vegetation, developed land, bare land, and agricultural land) in area, to which all lakes ($\geq 1 \text{ km}^2$) were converted in each subregion during the period of 1975–2015. The percentages of vegetation from Dongting and Poyang Lakes are highlighted in green circles. The horizontal lines (boxes and whiskers) in each boxplot are the 10th, 25th, 50th, 75th, and 90th percentiles, and the circles indicate the fifth and 95th percentiles.



Figure 7. (a) Distributions of lake classes in the MLY floodplain, i.e., Class I ("Urban"), Class II ("Connected"), and Class III ("Remainder"). (b) Boxplots show percentage distributions of land use, including the conversion of lakes in each class to aquaculture ponds, vegetation, developed areas, bare land, and cropland.

investigated the percentages of the land-use conversions of each lake in surface area (i.e., aquaculture ponds, vegetation, built-up areas, bare land, and cropland). The percentage distributions in each subregion are illustrated in Figure 6. For the whole floodplain, the transition from lakes to aquaculture ponds accounted for the highest proportion, on average (43.2%), followed by agricultural land (38.4%), vegetation (7.9%), builtup area (7.4%), and bare land (3.1%). The results demonstrate that local land reclamation and exploitation of lake resources were the leading factors driving the lake shrinkage in the MLY floodplain. Moreover, despite the dramatic exposure of wetland vegetation occurring in Dongting and Poyang Lakes (highlighted in green circles in Figure 6), Dongting Plain witnessed the largest transformation from lakes into aquaculture ponds and agricultural land (in terms of average proportion, 50.2% and 36.5%, respectively), and Poyang Plain had the highest percentage (68.4% on average) of conversion into cropland. In the other three subregions (i.e., Jianghan, Wanzhong, and the Yangtze Delta), the leading land-use transformations from lakes (in terms of average proportion) are consistent with the results

presented in Figure 4a. Additionally, under the accelerated process of urbanization, an increasing demand for land use in urban areas triggered the serious transformation from lakes into developed land, accounting for 7.4% of lake conversion, on average.

Driving Mechanisms among Different Lake Classes. In general, the lakes close to urban areas were vulnerable to the expansion of lakeside construction, e.g., real estate, traffic infrastructure, and tourist facility.¹⁵ These lakes are referred to as urban lakes in this study (Figure 7a), which are categorized as Class I ("Urban"), consisting of 16.2% (63 in number) of the total lakes in the floodplain. The lakes freely connected to the Yangtze River are categorized as Class II ("Connected"), including five lakes, i.e., Poyang, Dongting, Shijiu, and two oxbow lakes (Heiwa and Yangtze Ancient Channels).²⁶ The remaining 82.5% of lakes (321 in number) in the floodplain are categorized as Class III ("Remainder"), as their free connections with the Yangtze River were blocked by artificial channel diversion and local floodgate controls. As an important component of urban ecological environment, urban lakes have

been suffering from the severe influence of rapid urbanization, which is one of the most pervasive anthropogenic land conversions, leading to an area decrease of 22.2 km² (23.0% on average) in Class I lakes devoted to human settlements in the past few decades (Figure 7b). For example, Wuhan, the largest mega-city with an estimated population of 10.6 million in Central China,⁴⁸ has undergone an accelerated urbanization and significant shrinkage in urban lakes, such as Dong Lake, Nan Lake, and Sha Lake, which were converted into built-up areas occupying 32.4%, 59.4%, and 70.8% of loss in the respective surface areas (Figure S2a,b). The lakes in Class II (e.g., Poyang and Dongting) experienced a considerable exposure of wetland vegetation with a total area of 816.0 km² (51.0% on average), induced by the Yangtze flows. The pattern of change in Class III lakes was dominantly driven by extensive human activities for cropland and fish ponds, accounting for 45.7% and 39.0% of the decline in surface area, respectively. For instance, the shrinkage of the representative lakes, e.g., Tai Lake, Chao Lake, Hong Lake, and Liangzi Lake, was largely attributed to land reclamation for agricultural use or the enclosure practice for aquaculture (Figure S2a,b).

Ecological Impacts of Lake Degradation. Under the context of China's reform and opening-up policy since 1978, which has accelerated the process of industrialization and urbanization, excessive exploitation caused significant shrinkage of lakes during the period of 1975–2000. The Yangtze floodplain, one of the most fragile ecological systems in the world, has suffered from severe ecological consequences, e.g., increasing flood disasters, decreasing biodiversity, and environmental pollution. The human-induced rapid reduction in lake area across the MLY floodplain has a significant impact on the hydrological and ecological services.¹⁷

Flood Disasters. Numerous lakes across the MLY floodplain have played a critical role in accommodating the Yangtze floods. However, rapid shrinkage of the lake water area and their reduced connection with the Yangtze River has led to a significant decrease of the flood storage and drainage capacity.⁴⁴ For example, extensive reclamation has greatly reduced the water storage capacity of Dongting Lake from 2.68×10^{10} m³ in 1954 to 1.74×10^{10} m³ in 1983 (decreased by 35.1%),⁵⁰ which has increased the risk of flood hazard in the Yangtze River. From the 1950s to the 1990s, the recurrence frequencies of flood disasters on the Yangtze River have increased rapidly (Figure S3a).⁴⁹ In the 1950s, the recurrence interval of floods was as long as 10 years on average, but the average interval declined to 3-5 years between the 1960s and the 1980s and to only 2.5 years during the 1990s. In particularly, the 1998 flood, one of the most serious natural disasters on the Yangtze, was due not only to an extreme climatic catastrophe but also to the human-induced reclamation, deforestation, and soil erosion, causing enormous loss of economy amounted to 166 billion yuans or 20 billion U.S. dollars.⁵¹

To investigate the trend in the magnitude of Yangtze floods, water level data from the 1950s to the 1990s were collected from Duchang hydrological gauging station in the middle Yangtze.⁵² From the 1950s to the 1990s, the magnitude of the annual peak water level recorded at the Duchang station has gradually increased (Figure S3b). The annual maximum water level increased from 18.51 m in the 1950s to 20.44 m in the 1990s. In addition, there has been a significant increase in the frequency of the peak water level, suggesting increasing risks of severe floods.⁵³ During the 1950s, water level exceeded 19 m in only two years, while in the 1990s, water level exceeded 19 m

almost every year and exceeded 20 m in five years. The data analyzed in this study have all shown a notable increase in the frequency and magnitude of floods after the 1950s, in parallel with the rapid decline of lake water area.

Aquatic Plants and Fish. The lakes in the Yangtze Basin, as well as the Yangtze and its tributaries, provide important habitats for aquatic animals and plants. Nevertheless, severe degradation of lakes and wetlands over the past several decades has resulted in rapid loss of habitats, inducing a decline of biodiversity and extinction of endemic species. For example, the species richness of aquatic plants has shown a significant decrease during the past 50 years in Futou, Hong, and Dong lakes of middle Yangtze. The number of plant species in Futou Lake decreased from 70 to 61 between the early 1980s and the 2000s, i.e., nine plant species disappeared from the lake.⁵⁴ At Hong Lake, the aquatic plant richness has declined from 102 to 94 during the past few decades, as a result of the excessive development of fisheries. Since the 1970s, the dominant plant species in Dong Lake, such as sago pondweeds, duck lettuce, and ivy-leaf duckweed, have mostly disappeared, mainly due to the overexploitation of lake resources.⁵

Moreover, the loss of fish species was also increasing at an alarming rate. The species richness of fishes in Tai, Liangzi, and Dong lakes has dropped rapidly in the recent decades, consistent with the decline of lake water area. Before the 1980s, there were over 107 species of fish in Tai Lake, but only 60 species of fish could be collected in the 2000s.⁵⁶ The number of fish species in Liangzi Lake declined from 75 to 54 during the 1970s and 1980s. Similarly, Dong Lake experienced a rapid loss of the fish species (from 67 to 38) between the 1960s and 1990s, including some rare species such as the Yangtze grenadier anchovy and Reeve's shad.^{21,55} The Baiji dolphin, a well-known endemic species of the Yangtze River, decreased dramatically from 6000 in the 1950s to 60 in 1998, and now this species may become extinct in this study area.²¹

Water Quality. Excessive exploitation of lake resources for agricultural, aquaculture, or urban purposes has also caused the severe deterioration of water quality of the lakes across the MLY floodplain.⁵⁷ For example, the water surface of East Tai Lake was occupied rapidly by high-intensity enclosure culture since the 1990s, leading to significant increases in nutrient loading to the lake, such as nitrogen and phosphorus. From 1960 to 1981, total nitrogen (TN) increased dramatically from 0.05 to 0.9 mg/L, chemical oxygen demand (COD_{Mn}) increased by 49% (from 1.9 to 2.83 mg/L), and total phosphorus (TP) remained stable.⁵⁷ Since the 1990s, Tai Lake has reached the eutrophic state, with TN, TP, and COD_{Mn} increased from 2.349, 0.058, and 3.9 mg/L in 1990 to 2.7, 0.133, and 6.33 mg/L in 2000, respectively.⁵⁷

In addition, rapid urbanization and industrialization has resulted in land-use alterations in the watershed area of urban lakes, increasing the rate of urban contaminant inputs to the lakes.^{58,59} The conversions of agricultural land in the watershed area to urban areas may induce even greater levels of water pollution.⁶⁰ For example, in the watershed of Tangxun Lake, TN and TP loads have showed an increasing tendency since 1991, and the high-value areas of pollutant loads have expanded from north to south with the increase of developed areas.⁶¹ In Dong Lake, TN and COD_{Mn} increased from 1.92 and 4.7 mg/L in 1986 to 2.46 and 7.6 mg/L in 2000, respectively.⁵⁷ Water pollution caused by rapid urbanization has profound effects on the function of the lake as a drinking water supply, and thus exacerbated the lack of accessible drinking water in the cities.⁶²

Environmental Science & Technology

To mitigate these negative impacts, local governments have implemented a series of wetland restoration projects, such as establishing national wetland reserves, prohibiting impoldering and returning cropland to lakes, which contributed to the slight increase in surface area of lakes after the year 2000.^{15,63} For instance, Liangzi Lake, with the largest water storage capacity in Hubei Province, experienced an area shrinkage from 350.0 km² in 1975 to 286.2 km² in 2000 (a decrease of 18.2%), mainly due to the excessive lakeside enclosure for fish rearing. The implementation of lake conservation and restoration policies, e.g., prohibiting the enclosure practice and removing barriers in fish ponds, was conducive to an increase of 2.8% in lake area (from 286.2 to 304.8 km²) during 2000-2015. However, the lake areas of each subregion are forecast to decrease until the year 2020 (Figure S4) except for the Jianghan Plain (with a slight increase). For the whole MLY floodplain, the total area of lakes ($\geq 10 \text{ km}^2$) is expected to experience a reduction of 2.58% from 2015 to 2020. Under the accelerated urbanization and population growth, the numerous valuable lakes across the Yangtze floodplain will continue to suffer from land-use alternations for agriculture, aquaculture, and human settlements, with the ongoing vulnerability to both human activities and climate change.^{5,15,26} More effective action is urgently required to balance the trade-offs between human consumption and ecological conservation, and the further reconstruction of natural wetland ecosystems will be a tremendous challenge for policymakers.

ASSOCIATED CONTENT

S Supporting Information

The Supporting Information is available free of charge on the ACS Publications website at DOI: 10.1021/acs.est.6b04260.

Tables showing descriptions of Ramsar sites, acquisition dates of the Landsat images, changes in the surface area of lakes, validation results of Landsat water areas, and accuracy assessments for land-use changes. Figures showing examples of lake conversions in the MLY floodplain, charts indicating area and proportion of landuse types, frequency and magnitude of floods on the Yangtze River, a regression forecast, examples of the interpretation of Landsat images, the geographic distribution of validation sample points and error, and comparisons of Landsat mapped lake areas and reference datasets. Additional details on uncertainty analysis(PDF)

AUTHOR INFORMATION

Corresponding Author

*Phone: +86-02768771318; fax: +86-02768778086; e-mail: xhuang@whu.edu.cn.

ORCID 0

Xin Huang: 0000-0002-5625-0338

Notes

The authors declare no competing financial interest.

ACKNOWLEDGMENTS

We thank the four anonymous reviewers for their insightful comments. This work was supported in part by the China National Science Fund for Excellent Young Scholars under grant no. 41522110, in part by the National Key Research and Development Program of China under grant no. 2016YFB0501403 and in part by the Foundation for the Author of National Excellent Doctoral Dissertation, China, under grant no. 201348.

REFERENCES

(1) Tranvik, L. J.; Downing, J. A.; Cotner, J. B.; Loiselle, S. A.; Striegl, R. G.; Ballatore, T. J.; Dillon, P.; Finlay, K.; Fortino, K.; Knoll, L. B.; Kortelainen, P. L.; Kutser, T.; Larsen, S.; Laurion, I.; Leech, D. M.; McCallister, S. L.; McKnight, D. M.; Melack, J. M.; Overholt, E.; Porter, J. A.; Prairie, Y.; Renwick, W. H.; Roland, F.; Sherman, B. S.; Schindler, D. W.; Sobek, S.; Tremblay, A.; Vanni, M. J.; Verschoor, A. M.; von Wachenfeldt, E.; Weyhenmeyer, G. A. Lakes and reservoirs as regulators of carbon cycling and climate. *Limnol. Oceanogr.* **2009**, *54*, 2298–2314.

(2) Smith, L. C.; Sheng, Y.; Macdonald, G. M.; Hinzman, L. D. Disappearing Arctic lakes. *Science* **2005**, 308 (5727), 1429–1429.

(3) Gao, H.; Bohn, T. J.; Podest, E.; Mcdonald, K. C.; Lettenmaier, D. P. On the causes of the shrinking of Lake Chad. *Environ. Res. Lett.* **2011**, *6* (3), 034021.

(4) Tao, S.; Fang, J.; Zhao, X.; Zhao, S.; Shen, H.; Hu, H.; Tang, Z.; Wang, Z.; Guo, Q. Rapid loss of lakes on the Mongolian Plateau. *Proc. Natl. Acad. Sci. U. S. A.* **2015**, *112* (7), 2281–6.

(5) Ma, R.; Duan, H.; Hu, C.; Feng, X.; Li, A.; Ju, W.; Jiang, J.; Yang, G. A half-century of changes in China's lakes: Global warming or human influence? *Geophys. Res. Lett.* **2010**, *37* (24), L24106.

(6) Liu, H.; Yin, Y.; Piao, S.; Zhao, F.; Engels, M.; Ciais, P. Disappearing lakes in semiarid Northern China: drivers and environmental impact. *Environ. Sci. Technol.* **2013**, 47 (21), 12107–12114.

(7) Sellinger, C. E.; Stow, C. A.; Lamon, E. C.; Qian, S. S. Recent Water Level Declines in the Lake Michigan-Huron System. *Environ. Sci. Technol.* **2008**, *42* (2), 367–373.

(8) Schindler, D. W.; Williamson, C. E.; Saros, J. E. G.; Vincent, W. F.; Smol, J. P. Lakes as sentinels and integrators for the effects of climate change on watersheds, airsheds, and landscapes. *Limnol. Oceanogr.* **2009**, *54*, 2349–2358.

(9) Adrian, R.; O'Reilly, C. M.; Zagarese, H.; Baines, S. B.; Hessen, D. O.; Keller, W.; Livingstone, D. M.; Sommaruga, R.; Straile, D.; Van Donk, E.; et al. Lakes as sentinels of climate change. *Limnol. Oceanogr.* **2009**, *54* (3), 2283–2297.

(10) Vörösmarty, C. J.; Green, P.; Salisbury, J.; Lammers, R. B. Global water resources: vulnerability from climate change and population growth. *Science* **2000**, *289* (5477), 284–288.

(11) Shiklomanov, I. A. Global water resources. *Nature Resources* **1990**, *9* (99), 34–43.

(12) Schwarzenbach, R. P.; Egli, T.; Hofstetter, T. B.; von Gunten, U.; Wehrli, B. Global water pollution and human health. *Annual Review of Environment & Resources* **2010**, 35 (1), 109–136.

(13) National Bureau of Statistics of China. *China Statistical Yearbook;* China Statistics Press: Beijing, China, 2011.

(14) World Wildlife Fund. List of Ecoregions. http://wwf.panda.org/ about_our_earth/ecoregions/ecoregion_list (accessed Jul 1, 2016).

(15) Cui, L.; Gao, C.; Zhao, X.; Ma, Q.; Zhang, M.; Li, W.; Song, H.; Wang, Y.; Li, S.; Zhang, Y. Dynamics of the lakes in the middle and lower reaches of the Yangtze River basin, China, since late nineteenth century. *Environ. Monit. Assess.* **2013**, *185* (5), 4005–4018.

(16) Du, Y.; Xue, H.-P.; Wu, S.-K.; Ling, F.; Xiao, F.; Wei, X.-H. Lake area changes in the middle Yangtze region of China over the 20th century. *J. Environ. Manage.* **2011**, *92* (4), 1248–1255.

(17) Fang, J.; Rao, S.; Zhao, S. Human-induced long-term changes in the lakes of the Jianghan Plain, Central Yangtze. *Front Ecol Environ* **2005**, 3 (4), 186–192.

(18) Zhao, S.; Fang, J.; Miao, S.; Gu, B.; Tao, S.; Peng, C.; Tang, Z. The 7-decade degradation of a large freshwater lake in central Yangtze River, China. *Environ. Sci. Technol.* **2005**, *39* (2), 431–6.

(19) Tao, X. L.; Bai, Y. P. The Spatial-Temporal Change Pattern of Wetland in the Middle-Lower Yangtze River: A Case Study of Wuhu, Anhui. *Appl. Mech. Mater.* **2014**, *513*, 3228–3232.

(20) Zeng, H.; Wu, J. Heavy metal pollution of lakes along the midlower reaches of the Yangtze River in China: intensity, sources and

Environmental Science & Technology

spatial patterns. Int. J. Environ. Res. Public Health 2013, 10 (3), 793–807.

(21) Fang, J.; Wang, Z.; Zhao, S.; Li, Y.; Tang, Z.; Yu, D.; Ni, L.; Liu, H.; Xie, P.; Da, L.; et al. Biodiversity changes in the lakes of the Central Yangtze. *Front Ecol Environ* **2006**, *4* (7), 369–377.

(22) Feng, L.; Hu, C.; Chen, X.; Cai, X.; Tian, L.; Gan, W. Assessment of inundation changes of Poyang Lake using MODIS observations between 2000 and 2010. *Remote Sensing of Environment* **2012**, *121* (2), 80–92.

(23) Yuan, Y.; Zeng, G.; Liang, J.; Huang, L.; Hua, S.; Li, F.; Zhu, Y.; Wu, H.; Liu, J.; He, X.; et al. Variation of water level in Dongting Lake over a 50-year period: Implications for the impacts of anthropogenic and climatic factors. *J. Hydrol.* **2015**, *525* (10), 450–456.

(24) Chen, X.; Yang, X.; Dong, X.; Liu, E. Environmental changes in Chaohu Lake (southeast, China) since the mid 20th century: The interactive impacts of nutrients, hydrology and climate. *Limnologica* **2013**, 43 (1), 10–17.

(25) Guo, L. Ecology. Doing battle with the green monster of Taihu Lake. *Science* **2007**, 317 (5842), 1166–1166.

(26) Wang, J.; Sheng, Y.; Tong, T. S. D. Monitoring decadal lake dynamics across the Yangtze Basin downstream of Three Gorges Dam. *Remote Sensing of Environment* **2014**, *152*, 251–269.

(27) Du, Y.; Xue, H.-p.; Wu, S.-j.; Ling, F.; Xiao, F.; Wei, X.-h. Lake area changes in the middle Yangtze region of China over the 20th century. *J. Environ. Manage.* **2011**, *92* (4), 1248–1255.

(28) Ramsar Convention Bureau. The List of Wetlands of International Importance Website. http://www.ramsar.org/pdf/sitelist.pdf (accessed Jul 1, 2016).

(29) Hubei Bureau of Statistics. *Hubei Statistical Yearbook;* China Statistics Press: Beijing, China, 1985–2015.

(30) Hunan Bureau of Statistics. *Hunan Statistical Yearbook*; China Statistics Press: Beijing, China, 1985–2015.

(31) Jiangxi Bureau of Statistics. *Jiangxi Statistical Yearbook*; China Statistics Press: Beijing, China, 1985–2015.

(32) Anhui Bureau of Statistics. *Anhui Statistical Yearbook*; China Statistics Press: Beijing, China, 1985–2015.

(33) Jiangsu Bureau of Statistics. *Jiangsu Statistical Yearbook*; China Statistics Press: Beijing, China, 1985–2015.

(34) McFeeters, S. K. The use of the Normalized Difference Water Index (NDWI) in the delineation of open water features. *International Journal of Remote Sensing* **1996**, *17* (7), 1425–1432.

(35) Xu, H. Modification of normalised difference water index (NDWI) to enhance open water features in remotely sensed imagery. *International Journal of Remote Sensing* **2006**, *27* (14), 3025–3033.

(36) Tucker, C. J. Red and photographic infrared linear combinations for monitoring vegetation. *Remote Sensing of Environment* **1979**, 8 (2), 127–150.

(37) Zha, Y.; Gao, J.; Ni, S. Use of normalized difference built-up index in automatically mapping urban areas from TM imagery. *International Journal of Remote Sensing* **2003**, *24* (3), 583–594.

(38) Chen, W.; Liu, L.; Zhang, C.; Wang, J.; Wang, J.; Pan, Y. Monitoring the seasonal bare soil areas in Beijing using multitemporal TM images; Proceedings of the IEEE International Geoscience and Remote Sensing Symposium: Anchorage, Alaska, 2004; pp 3379–3382.

(39) Yuan, F.; Sawaya, K. E.; Loeffelholz, B. C.; Bauer, M. E. Land cover classification and change analysis of the Twin Cities (Minnesota) Metropolitan Area by multitemporal Landsat remote sensing. *Remote Sensing of Environment* **2005**, *98* (2–3), 317–328.

(40) Google Earth. Home Page. http://earth.google.com (accessed Sep 1, 2016).

(41) Ma, R. H.; Wang, S. M.; Li, A. N.; Wu, J. L.; Yang, G. S.; Duan, H. T.; Jiang, J. H.; Feng, X. Z.; Kong, F. X.; Xue, B.; et al. China's lakes at present:Number, area and spatial distribution. *Sci. China: Earth Sci.* **2011**, *54* (2), 283–289.

(42) Yang, S. W.; Dong, B.; Liu, L.; Sun, L.; Sheng, S. W.; Wang, Q.; Peng, W.; Wang, X.; Zhang, Z.; Zhao, J. Research on Vegetation Coverage Change in Sheng Jin Lake Wetland of Anhui Province. *Wetlands* **2015**, *35*, 1–6. (43) Lizhong, Y.; Yu, X.; Shuming, C.; XiaoYang, Z. A GIS-based study on recent environmental change in Honghu Lake. *Hupo Kexue* **1993**, *5*, 350–357.

(44) Feng, L.; Han, X.; Hu, C.; Chen, X. Four decades of wetland changes of the largest freshwater lake in China: Possible linkage to the Three Gorges Dam? *Remote Sensing of Environment* **2016**, *176*, 43–55.

(45) Mei, X.; Dai, Z.; Du, J.; Chen, J. Linkage between Three Gorges Dam impacts and the dramatic recessions in China's largest freshwater lake, Poyang Lake. *Sci. Rep.* **2016**, *5*, 18197.

(46) Feng, L.; Hu, C.; Chen, X.; Zhao, X. Dramatic Inundation Changes of China's Two Largest Freshwater Lakes Linked to the Three Gorges Dam. *Environ. Sci. Technol.* **2013**, *47* (17), 9628–9634. (47) Nilsson, C.; Revenga, C.; Reidy, C. A.; Dynesius, M.

Fragmentation and flow regulation of the world's large river systems. *Science* **2005**, 308 (5720), 405–408.

(48) Wuhan Bureau of Statistics, *Wuhan Statistical Yearbook*; China Statistics Press: Beijing, China, 2015.

(49) Yin, H.; Li, C. Human impact on floods and flood disasters on the Yangtze River. *Geomorphology* **2001**, *41* (2–3), 105–109.

(50) Li, Y.; Deng, J.; Sun, Z.; Li, R. Sediment deposition and variation of flood storage capacity in Dongting Lake. *Journal of Hydraulic Engineering* **2000**, 31 (12), 48–52.

(51) Zong, Y.; Chen, X. The 1998 Flood on the Yangtze, China. Natural Hazards 2000, 22 (2), 165–184.

(52) Changjiang Water Conservancy Committee, Atlas of the Changjiang River Flood Prevention; Science Press: Beijing, China, 2001.

(53) Yu, F.; Chen, Z.; Ren, X.; Yang, G., Analysis of historical floods on the Yangtze River, Characteristics and explanations, China. *Geomorphology* 2009, *113* (3–4), 210–216.

(54) Peng, Y.; Jian, Y.; Wang, J.; Ni, L. A comparative study on aquatic plant diversity in five largest lakes of Hubei Province in China. *Acta Hydrobiologica Sinica* **2004**, *28* (5), 464–470.

(55) Liu, J. Ecological studies on Donghu Lake (II); Science Press: Beijing, China, 1995.

(56) Songquan, Z.; Zhengwen, L.; Xiaohong, Gu Changes of the fish fauna and fish yield analysis in Lake Taihu. *Hupo Kexue* **2007**, *19* (6), 664–669.

(57) Cheng, X. Y.; Li, S. J. An analysis on the evolvement processes of lake eutrophication and their characteristics of the typical lakes in the middle and lower reaches of Yangtze River. *Chin. Sci. Bull.* **2006**, *51* (13), 1603–1613.

(58) Diamond, M. L.; Hodge, E. Urban Contaminant Dynamics: From Source to Effect. *Environ. Sci. Technol.* **2007**, *41* (11), 3796–3800.

(59) Melymuk, L.; Robson, M.; Csiszar, S. A.; Helm, P. A.; Kaltenecker, G.; Backus, S.; Bradley, L.; Gilbert, B.; Blanchard, P.; Jantunen, L.; Diamond, M. L. From the City to the Lake: Loadings of PCBs, PBDEs, PAHs and PCMs from Toronto to Lake Ontario. *Environ. Sci. Technol.* **2014**, *48* (7), 3732–3741.

(60) Chao, G.; Zhu, J. Y.; Dai, K. W.; Song, G.; Dou, Y. J. Impact of Rapid Urbanization on Water Quality and Related Mitigation Options in Taihu Lake Area. *Scientia Geographica Sinica* **2003**, 23 (6), 746–750.

(61) Yanhua, Z.; Song, H.; Wenting, Z.; Hongyan, L.; Qinghui, Z.; Thuminh, N.; Beibei, N.; Wanyi, L. Simulation of the spatial and temporal changes of complex non-point source loads in a lake watershed of central China. *Water Sci. Technol.* **2013**, 67 (9), 2050–2058.

(62) Shao, M.; Li, W.; Tang, X.; Zhang, Y. City clusters in China: air and surface water pollution. *Frontiers in Ecology & the Environment* **2006**, *4* (7), 353–361.

(63) Chengcheng, W.; Hongzhi, W.; Ablat, X. Spatial and Temporal Changes of Lake Wetlands in Jianghan Plain After the Implementing of 'Thirty-Six-Word Policy'. *Procedia Environ. Sci.* **2011**, *10*, 2574–2580.

3677